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Research Article

Direct evidence of the impact of early-life exposure to ambient PM_{2.5} air pollution on later-childhood height-for-age in India

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17 **The disease burden from ambient fine particulate (PM_{2.5}) exposure in India has been**
18 **estimated so far using risk functions based on studies done elsewhere. Here we provide the**
19 **first direct evidence of the impact of ambient PM_{2.5} exposure on child health in India using**
20 **measurements from nationally-representative anthropometric data matched to satellite-**
21 **based exposure data. We apply fixed effect regression with child height-for-age as the**
22 **dependent variable and district-month-level early-life exposure to ambient PM_{2.5} as the**
23 **independent variable. We show that a 100 $\mu\text{g}/\text{m}^3$ increase in ambient PM_{2.5} exposure leads**
24 **to a 0.05 standard deviation decrease in height-for-age after controlling for district-specific**
25 **seasonality, household properties, and other confounding factors. We find effects on both**
26 **rural and urban children, and cannot reject that the shape of the concentration-response**
27 **curve is linear. Because average exposure to ambient particulate pollution is high in India,**
28 **our results recommend ambient air pollution as public health policy priority.**

29 Ambient and household PM_{2.5} exposure have been linked causally to various child health outcomes
30 such as lower respiratory infection (LRI)¹, sudden infant death syndrome², low birth weight^{3,4},
31 intrauterine growth retardation^{5,6} and reduced size⁷. The recent Disease Burden of India (DBI)
32 study⁸ attributes 5.1 (4.1-6.3) million disability adjusted life years (DALY) and 0.06 (0.04-0.07)
33 million deaths of children (<5 years) to LRI, due to ambient PM_{2.5} exposure in India. The child
34 mortality burden due to household PM_{2.5} exposure in India is estimated to be equally large at 0.05
35 (0.03-0.06) million. However, these estimates relied on exposure-risk functions that were
36 developed from epidemiological studies carried out in developed countries. Direct evidence of the
37 impact of ambient PM_{2.5} exposure on child health in India is lacking so far.

38 One widely-studied marker of early-life health insults in India is the average height of children⁹.
39 Children in India are unusually short in international comparison, on average. Many correlates and
40 causes of India's mean child height deficit (or stunting), a clinically extreme height deficit, have
41 been documented in the demographic, epidemiological, and econometric literatures¹⁰⁻¹². Intra-
42 household exposure to particulate matter has also been found to impact child growth. Prevalence
43 of stunting was significantly higher (relative risk ratio RRR = 1.84, 95% uncertainty interval (UI)
44 1.44-2.36) amongst children living in the households that use solid fuel compared to the children
45 living in the households using clean fuel⁵. Using data from the 2005-2006 National Family Health
46 Survey (NFHS), a recent study⁶ has shown strong evidence that household solid fuel exposure
47 increases the risk of stunting and reduces the height-for-age measure of the Indian children. More
48 recently, ambient air pollution is also identified as a factor to impact child growth. Another study⁷
49 observed significant increases in the relative risk of child stunting and wasting in Bangladesh
50 associated with higher levels of in utero exposure to ambient air pollution. Ambient PM_{2.5} exposure
51 in India is quite large and varies in the range 3.7 to 148 $\mu\text{g}/\text{m}^3$ at annual scale^{8,9}. To our knowledge,
52 no study has ever been carried out to examine the impact of early-age ambient PM_{2.5} exposure on
53 child height-for-age in India.

54 To address this question in a sample representative of the population of children under five in India
55 – a population exposed to a large range of ambient PM_{2.5} – here we report an observational analysis
56 of India's 2015-2016 Demographic and Health Survey (DHS), matched to district-month level air
57 pollution, as measured by satellite remote sensing. The association between child height-for-age

58 and early-life exposure to air pollution is estimated using a fixed effects econometric strategy that
59 accounts for fixed differences across villages, for secular trends over time, and for district-specific
60 seasonal patterns. The resulting association between childheight outcomes and exposure to
61 ambient $PM_{2.5}$ is identified off of unpredictable deviations from these trends. We show that
62 ambient $PM_{2.5}$ exposure reduces child growth in India across rural and urban areas. The rate of
63 decrease of child height-for-age with an increase in ambient $PM_{2.5}$ exposure at their month of births
64 is consistently similar across the seasons and other confounding factors. We also show evidence
65 for a linear shape of the exposure-response function. These results are the first direct confirmation
66 of the impact of early-age ambient $PM_{2.5}$ exposure on child growth in India (or anywhere in the
67 world). We hope that our results might help formulate policy to curb ambient $PM_{2.5}$ exposure at
68 regional scale.

69 **Results**

70 **Summary and descriptive statistics.** Figure 1 describes the study population and sample. Height
71 is measured for 225,002 children under 5 in the DHS. We are able to match air pollution data to
72 children born before January 2016, who are 97% of those with measured height, resulting in a final
73 sample of 218,152 children. Summary statistics for these children are presented in Table 1. As
74 Table 1 shows, the DHS is also a rich source of further information on children and their
75 households, which we exploit in robustness checks as regression controls. Table 1 presents sample
76 means that summarize our data. By separating the sample by quintiles of ambient $PM_{2.5}$ exposure,
77 the table describes the correlates of the independent variable, and therefore some potential omitted
78 variable bias threats. Children who are exposed to higher ambient $PM_{2.5}$ also tend to be
79 disadvantaged in other ways. They come from larger families, have shorter mothers, live in
80 households that are more likely to defecate in the open, use traditional fuel, and live in the poorer
81 northern plains states of Uttar Pradesh and Bihar. However, much of these correlations are
82 absorbed by our controls and by our primary sampling units (PSU; Figure 1) and seasonality fixed
83 effects.

84 **Non-parametric analysis.** Before proceeding to our main regression results, we use non-
85 parametric methods to illustrate the relationship between ambient $PM_{2.5}$ exposure and subsequent
86 child height. Figures 2, 3, and 4 use locally-weighted polynomial regression; all three show
87 evidence of a robust association. Figure 2 reveals a negative gradient between ambient $PM_{2.5}$
88 exposure and child height for both rural and urban children. Overall, rural children are shorter, on
89 average, because they are more exposed to other factors associated with growth faltering^{10,11}; this
90 is visible in the fact that the rural line is below the urban line. The principal results of Figure 2,
91 however, are that both lines have an apparently linear downwards gradient, and that they are
92 parallel, which is consistent with a comparable gradient for rural and urban children.

93 The seasonality of exposure to ambient $PM_{2.5}$ in India is reflected in Figure 3. Here, each season-
94 of-birth is plotted separately, to ensure that season of birth is not a biasing omitted variable.
95 Ambient $PM_{2.5}$ levels reach the highest levels in the winter months of November through January,
96 which is visible in the fact that this line extends the furthest to the right. However, Figure 3
97 provides evidence that the pattern in Figure 2 does not merely reflect a seasonal trend in height,
98 because similarly-steep downward gradients are present in all four seasons. In other words, the

99 season of birth may matter for a child's outcomes, but it is not a confounder in the gradient that
100 we document.

101 Finally, by splitting the sample by decile of mother's height, Figure 4 speaks to the possibility that
102 the gradient on Figure 2 merely reflects confounding heterogeneity across household environments
103 or among children's genetic endowments. Each line is separately computed for children of mothers
104 with similar heights. The lines are approximately parallel: within each decile, children exposed to
105 more ambient PM_{2.5} in their month of birth are shorter, on average. Mothers' height (and its
106 correlates) does not appear to be a potentially biasing omitted variable for our results.

107 **Effect of month-of-birth exposure.** Table 2 presents our main results: fixed effect regression
108 results following equation 1 (see Methods for more details). Ambient PM_{2.5} exposure is divided
109 by 100 for ease of interpretation of the coefficients. Across the alternative specifications in
110 columns 1 through 5, a 100 $\mu\text{g}/\text{m}^3$ increase in ambient PM_{2.5} exposure is associated with an
111 approximately 0.05 standard deviation decrease in child height-for-age. Column 4 verifies that the
112 result is unchanged after controlling for household fuel use. Column 5 is a falsification test.
113 Ambient PM_{2.5} exposure two years before the child is born does not predict height and does not
114 change the coefficient of interest. Columns 7 and 8 find similar results when state-month or PSU-
115 month fixed effects for seasonality are used instead of district-months.

116 Although not reported in the table, we conducted a further robustness check that our result is not
117 driven by the January height pattern documented in the DHS literature. Omitting children born in
118 January results in an essentially unchanged estimate of -0.052 (standard error = 0.023; $p = 0.024$).
119 Additionally replacing year of birth fixed effects with a larger set of state-specific year of birth
120 fixed effects reduces precision by consuming degrees of freedom, but does not qualitatively change
121 the estimate (-0.069, standard error = 0.030, $p = 0.022$ in the most fully controlled specification).

122 Figure 5 presents seven estimates of equation 2 (see the Methods section), each for ambient PM_{2.5}
123 exposure in a separate three-month age range. Only the early-life period at and immediately after
124 birth shows a coefficient that is statistically distinguishable from zero. Most of the other
125 coefficients are close to zero, and none is as large in absolute value as the one for ages 0-2 months.
126 These results are consistent with evidence in the literature¹² that early-life is a critical period for
127 the determination of child height. Some prior literature has documented evidence for effects on
128 health of *in utero* exposure¹³; although we do not detect *in utero* effects, the confidence intervals
129 on pre-birth exposure cannot rule out effects about half as large as the effect that we find for
130 exposure in the first months of life.

131 **Shape of concentration-response function.** Three tests for non-linear concentration-response
132 functions each fail to reject that a linear shape fits the data. Moreover, each approach suggests that,
133 if anything, effects are steeper at higher concentration levels. Table 2 includes non-linear candidate
134 function shapes. As column 6 shows, a natural log functional form – which would be consistent
135 with the hypothesis that the concentration-response function shows diminishing marginal costs –
136 has a coefficient that is statistically distinguishable from zero, but fits the data less well (as
137 measured by a t -statistic) than a linear form. Column 9 includes a linear spline that allows a
138 different slope at above-median levels of ambient PM_{2.5}. Although the two PM_{2.5} terms are jointly

139 statistically significant ($F = 2.99$; $p = 0.051$), neither is individually statistically significantly
140 different from zero. The coefficient on the spline term is negative, indicating that, although this
141 model does not fit the data better than a simple linear concentration-response function, the sign
142 suggests a steeper concentration-response function at higher levels of exposure.

143 Finally, Figures 6 and 7 present two further ways of investigating whether the concentration-
144 response function has evidence of a non-linear shape. Figure 6 investigates higher-order
145 polynomials, beyond linear. It presents results for quadratic, cubic, quartic, and quintic polynomial
146 forms. Of these, only the quadratic form is jointly statistically significant, and its fit does not
147 improve on a linear functional form. The coefficient estimates for higher-order polynomials
148 suggest, in each case, effects that are, if anything, steeper at higher levels of exposure. Figure 7
149 graphically presents the results of a Cox-Box transformation, a standard parameterization of a
150 curved relationship, detailed in the Methods section. It presents log-likelihoods for a range of
151 power transformations of $PM_{2.5}$, estimated with and without the full set of controls. The likelihood
152 of the model is maximized at or just above an exponent of 1, indicating that a linear model (or
153 perhaps one with slightly increasing marginal effects) best fits these data.

154 Discussion

155 This paper reports an ecological analysis of variation in remotely-sensed ambient $PM_{2.5}$ exposure
156 data at the district-month level. Ecological analysis is often used to generate hypotheses for further
157 investigation using more rigorous methods. In this instance, there are inherent limits to the possible
158 study design: ambient air pollution is an important topic of study, but it must vary at a geographic
159 level, and is not amenable to experimental manipulation. Although place, time, and season fixed
160 effects limit the role of residual confounding, we are unable to use an econometric design that
161 exploits a specific, known source of variation in $PM_{2.5}$ exposure, such as a policy change.
162 Nevertheless, measurement error could, in principle, be improved by a study that records child-
163 level exposure to ambient $PM_{2.5}$ with a system of mobile child-level personal monitors.

164 This is the first direct evidence of the ambient $PM_{2.5}$ exposure impact on child health at a country
165 level. Although child height has traditionally been interpreted as a measure of “malnutrition,” it is
166 increasingly recognized that anthropometric outcomes such as height reflect the totality of early-
167 life net available nutrition, including losses due to diseases¹⁴, and including growth effects of lung
168 function. Our data do not allow us to observe disease directly; indeed our health data reflect only
169 conditions at the time of the survey, and not during the critical period of the child’s birth. However,
170 mechanisms in the literature are consistent with the effect that we document⁷. For example, lung
171 function growth has been linked to children’s exposure to particulate matter¹⁵. More generally,
172 child growth is highly correlated, at the population level, with early-life mortality, which has been
173 interpreted as a consequence of the role of infectious disease. Mortality is correlated with average
174 child height because survivors’ growth is “scarred” by its early-life disease¹⁶. This could plausibly
175 include respiratory disease.

176 Many studies in the air pollution literature use mortality as a dependent variable, typically from
177 census or vital registration data. But India does not have a vital registration system, like many
178 other developing countries. As Setel and colleagues¹⁷ explains: “Most people in Africa and Asia

179 are born and die without leaving a trace in any legal record or official statistic.” Therefore, we
180 study height-for-age as a dependent variable, because it is a summary of early-life health that is a
181 continuous variable, and therefore offers high statistical power even in a survey sample, relative
182 to dichotomized outcomes such as stunting or infant mortality ¹⁸.

183 We document an effect of early-life exposure to ambient PM_{2.5} on subsequent height-for-age in
184 later childhood, using India’s most recent DHS, which measures the children under five years old
185 of a nationally representative sample of reproductive age women. The effect size we estimate is
186 plausibly small for any one child, but many children are exposed to it. Moreover, ambient PM_{2.5}
187 concentrations in India are higher than World Health Organization guideline. With an effect size
188 of 0.05 associated with a linear difference of 100, the average child in India is about 0.027 height-
189 for-age standard deviations shorter than he or she would be if exposed to very low levels of air
190 pollution, an effect multiplied by almost 30 million births per year. Because the ambient PM_{2.5}
191 exposure is projected to increase in India in near future under climate change scenarios ¹⁹, the
192 health burden that we quantify here could potentially increase unless appropriate policy is taken
193 to reduce air pollution throughout India. In particular, although policy conversations often focus
194 on Delhi (and, to a lesser extent, other big cities), we find results throughout India for rural and
195 urban children, suggesting that the policy challenges are significantly broader than is commonly
196 understood. Because child height has lasting consequences for human capital ^{12,20}, this is a problem
197 with ramifications throughout the Indian society and economy.

198 **Methods**

199 **India’s 2015-2016 Demographic and Health Survey.** The dependent variable and regression
200 controls are taken from India’s most recent Demographic and Health Survey (hereafter DHS; in
201 India, the DHS is also known as the National Family Health Survey). These data were collected
202 from a nationally-representative sample of women of reproductive age. In particular, the sample
203 was constructed to permit district-level estimates for all 640 districts in India at the time of the
204 2011 census. These data were collected between January 2015 and November 2016.

205 Our dependent variable of interest is a child’s height-for-age *z*-score, scaled according to the World
206 Health Organization 2006 reference population mean and standard deviation by sex and age-in-
207 months (WHO 2006). In the DHS, height is measured for children less than five years old at the
208 time of the survey. The sex and month of birth (e.g. August 2011) is also recorded for each child
209 with measured height.

210 **Air pollution data by district-month.** Each child is matched to average ambient PM_{2.5} exposure
211 in his or her district of residence, during the month in which he or she was born. This matching
212 implicitly assumes that the district where children live at the time of the survey is the same as the
213 district where children lived when they were born.

214 The absence of systematic ground-based PM_{2.5} measurements at desirable spatial resolution
215 prompted us to use satellite-derived PM_{2.5} for this study. We use the Multiangle Imaging
216 SpecroRadiometer (MISR) retrieved daily aerosol optical depth (AOD) V22 product at 17.6 km
217 spatial resolution to estimate the PM_{2.5} with the help of a spatially and temporally varying
218 conversion factor (η). η is derived from of GEOS-Chem chemical transport model simulations and

219 depends on aerosol vertical distribution, emission and meteorological factors like temperature,
220 relative humidity and precipitation. Details about the conversion factor η are discussed
221 elsewhere^{9,21,22}. MISR AOD product was earlier extensively evaluated for the Indian subcontinent
222 ²³. The satellite-retrieved PM_{2.5} was bias-corrected using coincident ground-based quality
223 controlled measurements following our earlier study^{8,9}. The district-level statistics are extracted
224 using the shape files of the district boundaries in ArcGIS. We generate a monthly PM_{2.5} exposure
225 database for 15 years (2001-2015), although because height is only measured in the DHS for
226 children under 5, no child in our sample was born before 2010.

227 **Fixed effects econometric strategy.** The central empirical strategy of this paper is fixed effects
228 regression, with child height-for-age as the dependent variable, and early-life district-month-level
229 exposure to ambient PM_{2.5} as the independent variable of interest. The DHS is a cross-sectional
230 survey that measured children under five at different ages; because age is predictably correlated
231 with height-for-age²⁴, each regression therefore controls for 119 age-in-months-by-sex indicators.
232 Each regression also controls for fixed effects for place, season, and year, as detailed in ‘Results’
233 section. We first present the econometric strategy for the main result followed by further
234 investigation of exposure at other ages and robustness checks in which we allow the concentration-
235 response function to take non-linear shapes. All analysis in the paper is computed with Stata 12.1.

236 **Non-parametric descriptive regression.** Exposure to ambient PM_{2.5} is not randomly allocated
237 across children. This fact raises the possibility that any apparent association between air pollution
238 and child outcomes could, in fact, reflect omitted variables such as seasonality of births or
239 geographic heterogeneity across India. Graphs of non-parametric regressions in split samples can
240 be a method to investigate whether the variable over which the sample is split confounds the
241 relationship of interest²⁵. In particular, in Figures 2, 3, and 4 we plot locally-weighted kernel
242 regressions, computed in Stata with the default Epanechnikov kernel function. Each figure uses,
243 as the independent variable, ambient PM_{2.5} in the district month of birth and, as the dependent
244 variable, height-for-age residuals after regression on 119 age-in-months-times-sex fixed effects,
245 to account for the fact that children were measured at different ages.

246 Figure 2 splits the sample by rural and urban. Although much of the discussion of air pollution in
247 India focuses on urban places, this permits us to see if there is a gradient for rural children. Figure
248 3 splits the sample by season of birth; within each season, only observations in the 5th to 95th
249 percentiles of PM_{2.5} exposure are plotted (because non-parametric statistics of this sort require an
250 adequate sample size). This split permits us to investigate whether any gradient is present within
251 seasons, and not merely a reflection of omitted seasonality. Finally, in recognition of the fact that
252 PM_{2.5} exposure is high in the poorer regions of India, where other factors also constrain child
253 growth, Figure 5 splits the sample into ten partitions, according to deciles of mothers’ height;
254 within each decile of mother’s height, only observations in the 5th to 95th percentiles of PM_{2.5}
255 exposure are plotted. If a gradient is visible within each decile, then it is evidence that the
256 association between air pollution exposure and child height is not fully driven by the omitted
257 variables that are correlated with mothers’ height.

258 **Main strategy.** Our main analysis estimates fixed effects regressions of the following form:

$$259 \quad h_{ipdmy} = \beta x_{dmy} + \alpha_{pd} + \gamma_{dm} + \delta_y + X_{ipdmy}\theta + \varepsilon_{ipdmy} \quad (1)$$

260 where i indexes individual children, p places (survey PSUs, such as urban blocks or rural villages),
 261 d districts, m calendar months (such as February), and y calendar years, such as 2008. The
 262 dependent variable, h , is child i 's height-for-age z -score. The independent variable of interest,
 263 x_{dmy} , is PM_{2.5} in district d in month m of year y , corresponding to child i 's birth month. Fixed
 264 effects are α_{pd} , 27,266 local places (PSUs); γ_{dm} , 7,679 categories of district-month (such as for
 265 Februarys in Sitapur district, or Aprils in Kanpur district); and δ_y , 6 calendar years, to capture any
 266 secular time trend. Child-level covariates X_{ipdmy} include the age-by-sex fixed effects and a vector
 267 of extended controls.

268 We add fixed effects and controls in stages to verify that the main effect estimate, $\hat{\beta}$, is robust to
 269 respecification. In particular, we first estimate the model without PSU fixed effects. PSU fixed
 270 effects would account for any fixed geographic heterogeneity such as market ²⁶, local open
 271 defecation ¹⁰, or the religious composition of the neighborhood²⁷. We then add a set of extended
 272 regression controls, intended to control for other known determinants of child height: the height
 273 of the child's mother, the child's birth order, the number of siblings born to its mother by the time
 274 of the survey, household open defecation, the caste and religion of the child's household ¹¹, the
 275 and child's mother's relationship to the head of the household. We further add indicators for the
 276 household's cooking fuel type, as a proxy for household PM_{2.5} exposure.

277 Our main specifications control for district-month fixed effects²⁸. This strategy allows each district
 278 to have any distinct seasonality pattern, and identifies effects off of deviations from each district's
 279 seasonal patterns. Fertility is known to be predicted by seasonal patterns ²⁹, but parents would not
 280 be able to make fertility decisions (nine months in advance) based on the realized deviation from
 281 seasonal trends. Therefore, predictable seasonality does not confound our results. We include
 282 robustness checks with coarser (state-month) and finer (PSU-month) controls for seasonality.

283 Finally, we conduct several additional robustness tests. As a falsification test, we control for
 284 ambient PM_{2.5} in the same district-month, two years before the month of birth; if our identification
 285 strategy is credible, this control should not predict height nor change our estimate. Few studies^{30,31}
 286 both document a pattern in DHS data of a drop in child height of children born in January, relative
 287 to children born in December; they note that this pattern could bias studies that identify effects on
 288 height from child season of birth. This is not what our work does. We control for seasonality and
 289 identify off of deviations from it, and we control non-parametrically for age-in-months-times-sex.
 290 However, to verify that this pattern is not a threat to our conclusion, we compute a robustness
 291 check omitting children reported to be born in January.

292 Standard errors are clustered by 640 districts, to permit arbitrary correlation of error terms over
 293 space and time within districts ³². DHS data include sampling weights; although we use weights
 294 for our summary statistics in Table 1, sampling weights are not appropriate for estimating
 295 relationships ³³, so we do not use weights in any of our regression results.

296 **Age of exposure.** Our main specification only investigates the effect of exposure to ambient PM_{2.5}
 297 in the month of birth. In an extension, we consider exposure at other ages. We average over three-
 298 month age ranges, from -9 to -7 months (first trimester of pregnancy) to 9 to 11 months (the last
 299 quarter of the first year of life). Average PM_{2.5} in each age of exposure is used as the independent
 300 variable in a separate regression:

$$301 \quad h_{ipdmy} = \beta \left(\frac{x_{dmy}^{+0} + x_{dmy}^{+1} + x_{dmy}^{+2}}{3} \right) + \alpha_{pd} + \gamma_{dm} + \delta_y + X_{ipdmy}\theta + \varepsilon_{ipdmy} \quad (2)$$

302 where indices and fixed effects are as in regression equation (1), but the controls X include only
 303 the age-in-months-by-sex indicators, and not the full set of extended controls. Coefficient
 304 estimates and 95% confidence intervals are presented in Figure 5.

305 **Robustness of shape of the concentration-response function.** The shape of the concentration-
 306 response function has been a focus of the air pollution literature, in light of its importance for
 307 policy responses (Pope et al. 2015). Although the prior literature has emphasized the possibility of
 308 diminishing marginal costs (such that the extra damages from extra exposure decline at higher
 309 levels of exposure), there is little well-identified evidence on exposure to PM_{2.5} at levels as high
 310 as in India during the period studied.

311 Therefore, we perform three robustness checks in which we allow the shape of the concentration-
 312 response function to differ from the linear form in equation (1):

$$313 \quad h_{ipdmy} = f(x_{dmy}) + \alpha_{pd} + \gamma_{dm} + \delta_y + X_{ipdmy}\theta + \varepsilon_{ipdmy}. \quad (3)$$

314 First, in Table 2, we substitute in the natural log of PM_{2.5} in one specification, and a linear spline
 315 at the median of PM_{2.5} in another. Then, in Figure 6, we allow polynomial shapes of the
 316 concentration-response curve, of degree 1 through 5. Finally, in Figure 7, we implement a Box-
 317 Cox power transformation of the form $f(x) = x^\lambda$, for coefficients λ in steps of 0.1 from 0.1 to 2.0.
 318 We implement each power transformation in a separate regression, and plot the resulting log-
 319 likelihoods. If likelihood is maximized near $\lambda = 1$, then this procedure would suggest that a linear
 320 concentration-response function fits these data.

321 **Acknowledgments.** SD acknowledges DST-FIST grant (SR/FST/ESII-016/2014) for upgradation
 322 of computing facility at IIT Delhi. MISR aerosol data are archived in NASA Langley Research
 323 Atmospheric science data Center.

324 **Code availability.** The codes used to arrive at the results depicted in this study can be available
 325 upon request to the corresponding author.

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405 **Author contributions.** D.S. and S.D. developed the idea, D.S. carried out the main analysis, and
 406 S.C. carried out the satellite data analysis to generate the exposure data. D.S. and S.D. wrote the
 407 paper with inputs from all the authors.

408 **Competing financial interests.** The authors declare they have no actual or potential competing
 409 financial interests.

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411 **Table 1.** Summary statistics describing height-for-age sample from India's 2015-16 DHS.

	PM _{2.5} district-month of birth quintiles					
	full sample	1	2	3	4	5
PM _{2.5} in birth month	54.9	15.3	30.1	45.7	65.2	118.2
height-for-age	-1.50	-1.35	-1.45	-1.52	-1.59	-1.60
age in months	30.7	31.9	31.2	30.7	30.4	29.1
girl	0.48	0.49	0.48	0.48	0.47	0.48
rural	0.72	0.67	0.71	0.72	0.74	0.76
mother's height (cm)	151.7	152.2	151.9	151.6	151.4	151.2
uses LPG	0.33	0.42	0.36	0.33	0.28	0.25
uses traditional fuel	0.63	0.53	0.60	0.63	0.68	0.72
open defecation	0.47	0.40	0.47	0.49	0.51	0.49
birth order	2.18	1.97	2.09	2.18	2.30	2.38
sibsize	2.46	2.23	2.37	2.46	2.59	2.67
in UP or Bihar	0.31	0.09	0.18	0.29	0.41	0.57
<i>n</i> (children under 5)	218,152	52,947	43,942	40,831	40,551	39,881

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413 *Note:* Each number, other than sample sizes in the bottom row, is a sample mean. Girl, rural, uses
 414 LPG, uses traditional fuel, open defecation and in Uttar Pradesh (UP) or Bihar are each indicators
 415 (1 or 0) for that property of the child or household. Sibsize is a measure of fertility: the number of
 416 children born to the child's mother at the time of the survey. Sample means and quintiles are
 417 computed with DHS sampling weights (which is why *n* is not constant across quintiles).

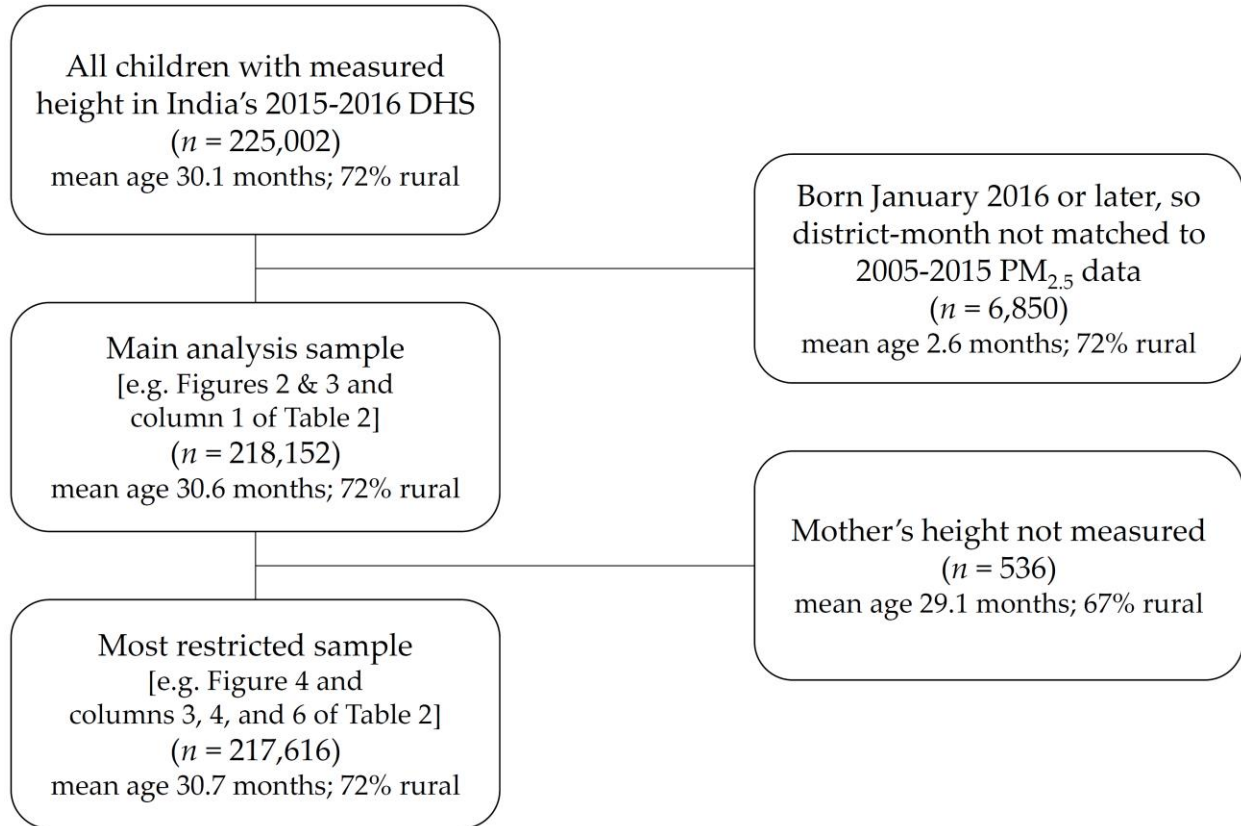
418 **Table 2.** Regression results: Child height-for-age z -score regressed on district-level $PM_{2.5}$ in month of birth, with fixed effects and
 419 covariate controls

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	(1)	(2)	(3)	(4)	(5)	(6)	(7)	(8)	(9)
$PM_{2.5} \div 100$	-0.0546**	-0.0494*	-0.0488*	-0.0477*	-0.0502*		-0.0352*	-0.0734*	-0.0286
	(0.0201)	(0.0210)	(0.0207)	(0.0206)	(0.0216)		(0.0178)	(0.0309)	(0.0555)
$PM_{2.5} \div 100$ 24 months earlier $\ln(PM_{2.5})$					-0.0107 (0.0209)				
						-0.0171 [†] (0.00987)			
mother's height (cm)			0.0470** (0.000932)	0.0467** (0.000925)	0.0467** (0.000925)	0.0467** (0.000925)			
$PM_{2.5} \div 100$ above median spline									-0.0281 (0.0668)
age-in-months \times sex	yes	yes	yes	yes	yes	yes	yes	yes	yes
district-month FEs	yes	yes	yes	yes	yes	yes			yes
year of birth FEs	yes	yes	yes	yes	yes	yes	yes	yes	yes
PSU FEs		yes	yes	yes	yes	yes	yes		yes
extended controls			yes	yes	yes	yes			
household fuel				yes	yes	yes			
state-month FEs							yes		
PSU-month FEs								yes	
n (children under 5)	218,152	217,285	216,745	216,745	216,745	216,745	217,286	115,586	217,285

421 *Note:* All columns present ordinary least squares fixed effects regressions with the child's height-for-age z -score as the dependent
 422 variable. FE = fixed effect; PSU = primary sampling unit (urban block or rural village). Standard errors clustered by 640 districts in
 423 parentheses. [†] $p < 0.10$; * $p < 0.05$; ** $p < 0.01$. In column 9, the spline variable is zero below the median $PM_{2.5}$ and is identical to
 424 $PM_{2.5}$ above the median. Sample sizes vary because some fixed effects categories lack within-category variation in the independent
 425 variable (resulting in that category being dropped), and because not all children's mothers' height was measured.

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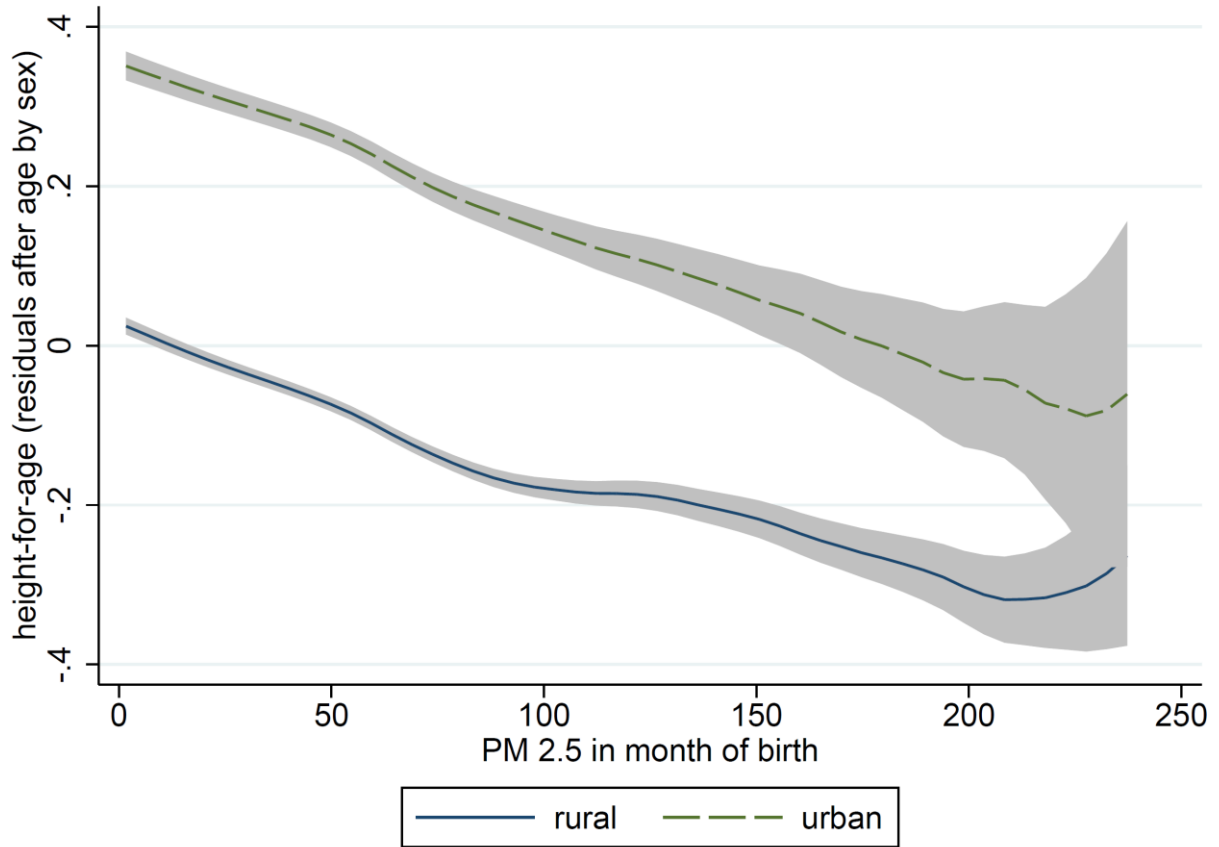


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Figure 1. Study sample with excluded or missing observations. In Table 2, some samples are smaller than 217,616 because fixed effects regression ignores categories within which there is no variation in the independent variable.

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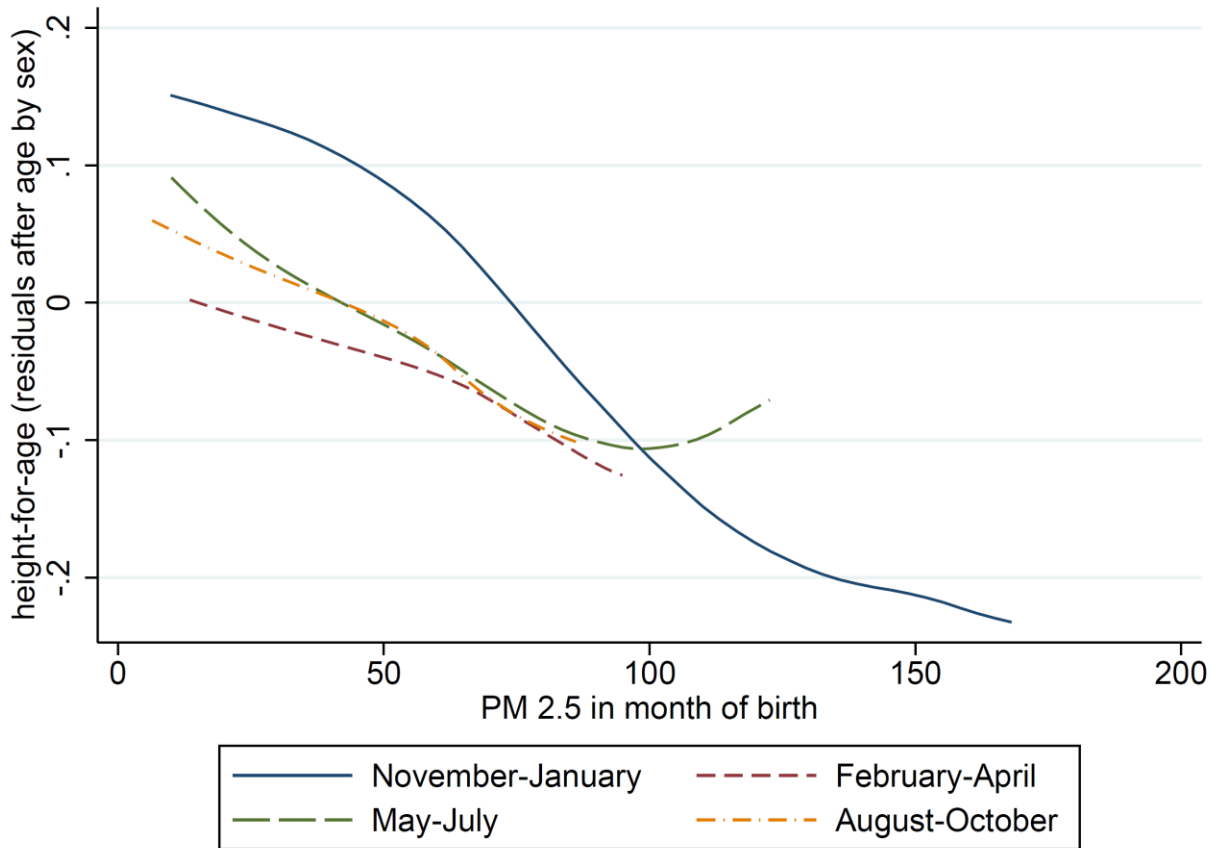


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441 **Figure 2.** Non-parametric association between child height and exposure to PM_{2.5} in the month of
442 birth. Curves are kernel-weighted local regressions. The vertical axis is the residual of child
443 height-for-age on 120 age-in-months by sex indicators.

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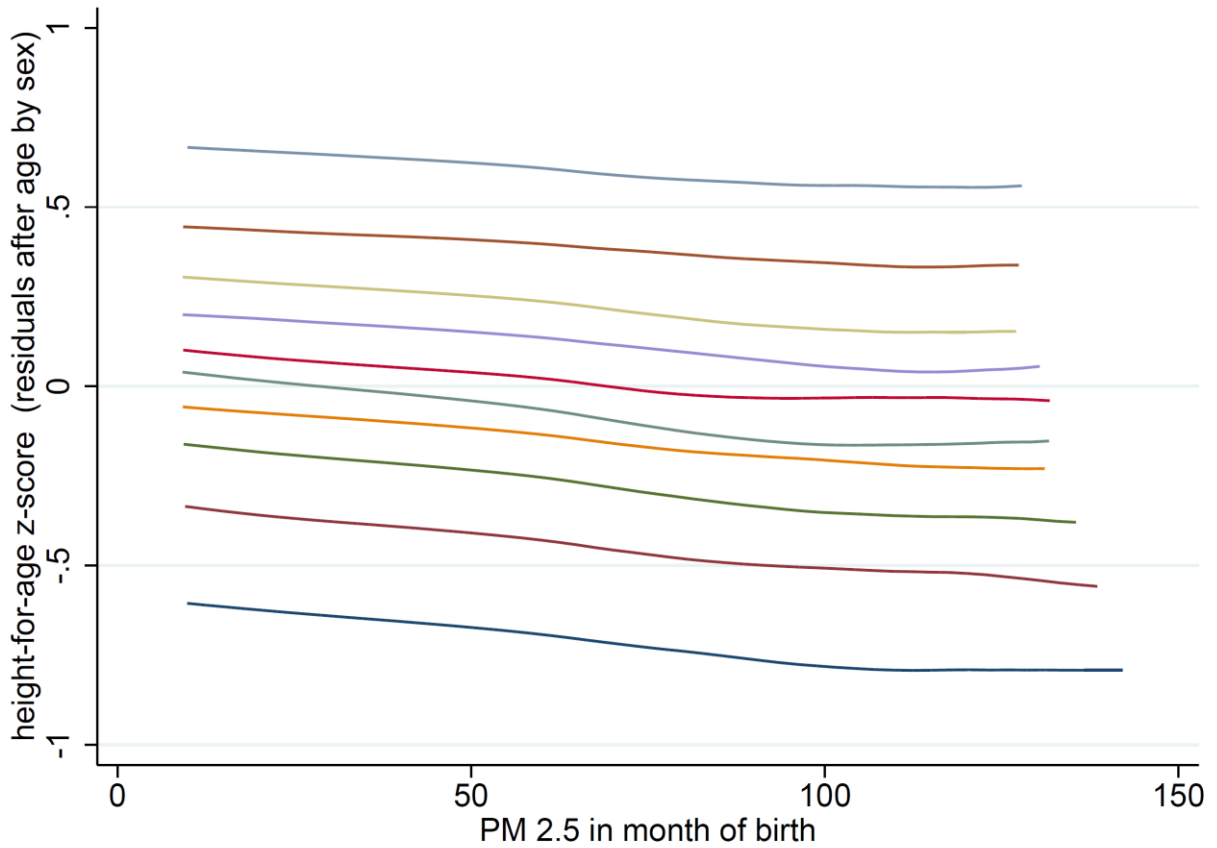
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450 **Figure 3.** Child height and exposure to PM_{2.5} in the month of birth, by season of birth. Curves are
451 non-parametric kernel-weighted local regressions. The vertical axis is the residual of child height-
452 for-age on 120 age-in-months by sex indicators.

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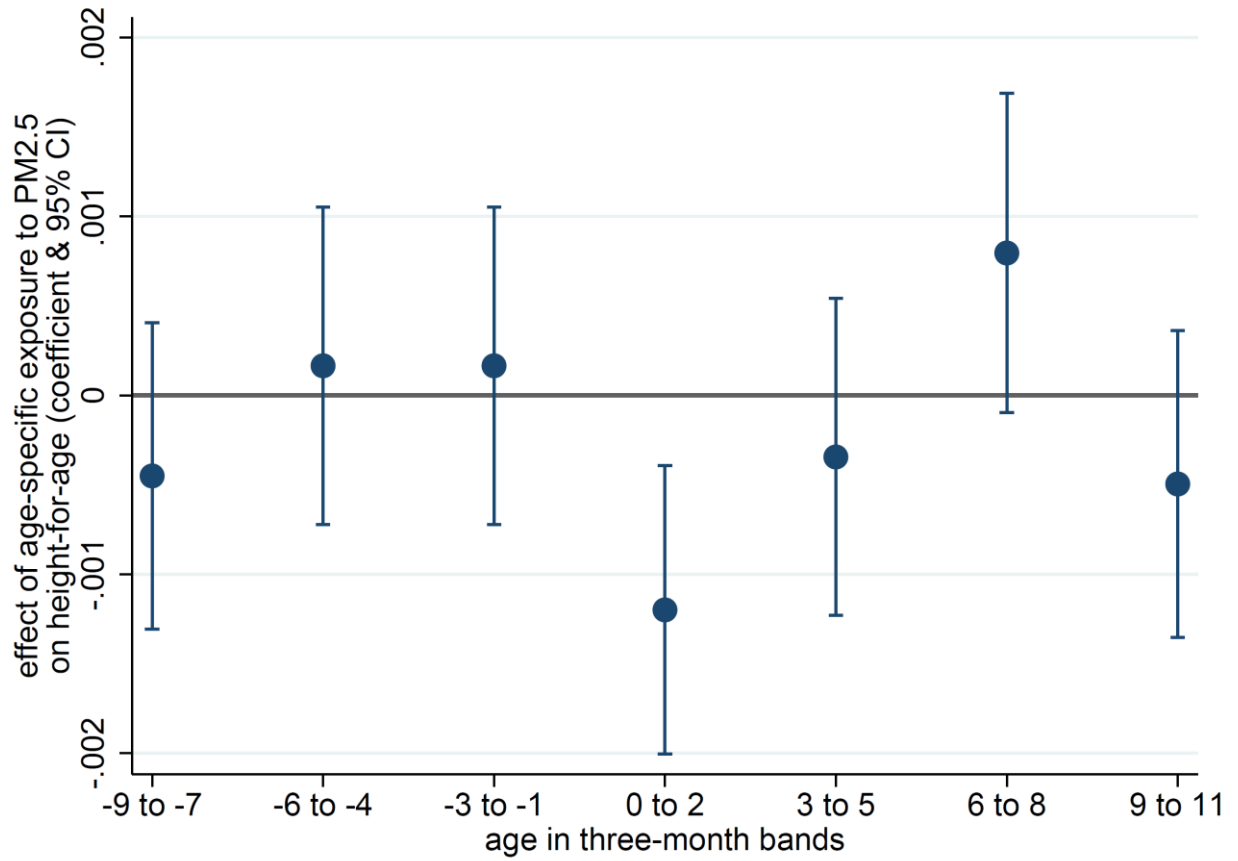
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459 **Figure 4.** Child height and exposure to PM_{2.5} in the month of birth, by decile of mother's height.
460 Curves are non-parametric kernel-weighted local regressions. Each separate curve includes only
461 children born to mothers in one of ten height deciles. The vertical axis is the residual of child
462 height-for-age on 120 age-in-months by sex indicators.

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468 **Figure 5.** Effects of PM_{2.5} exposure at various ages. Each dot is a coefficient and each range is a
469 95% confidence interval from a separate fixed effects regression of child height-for-age on the
470 average exposure to PM_{2.5} in the months, relative to birth, specified along the horizontal axis.

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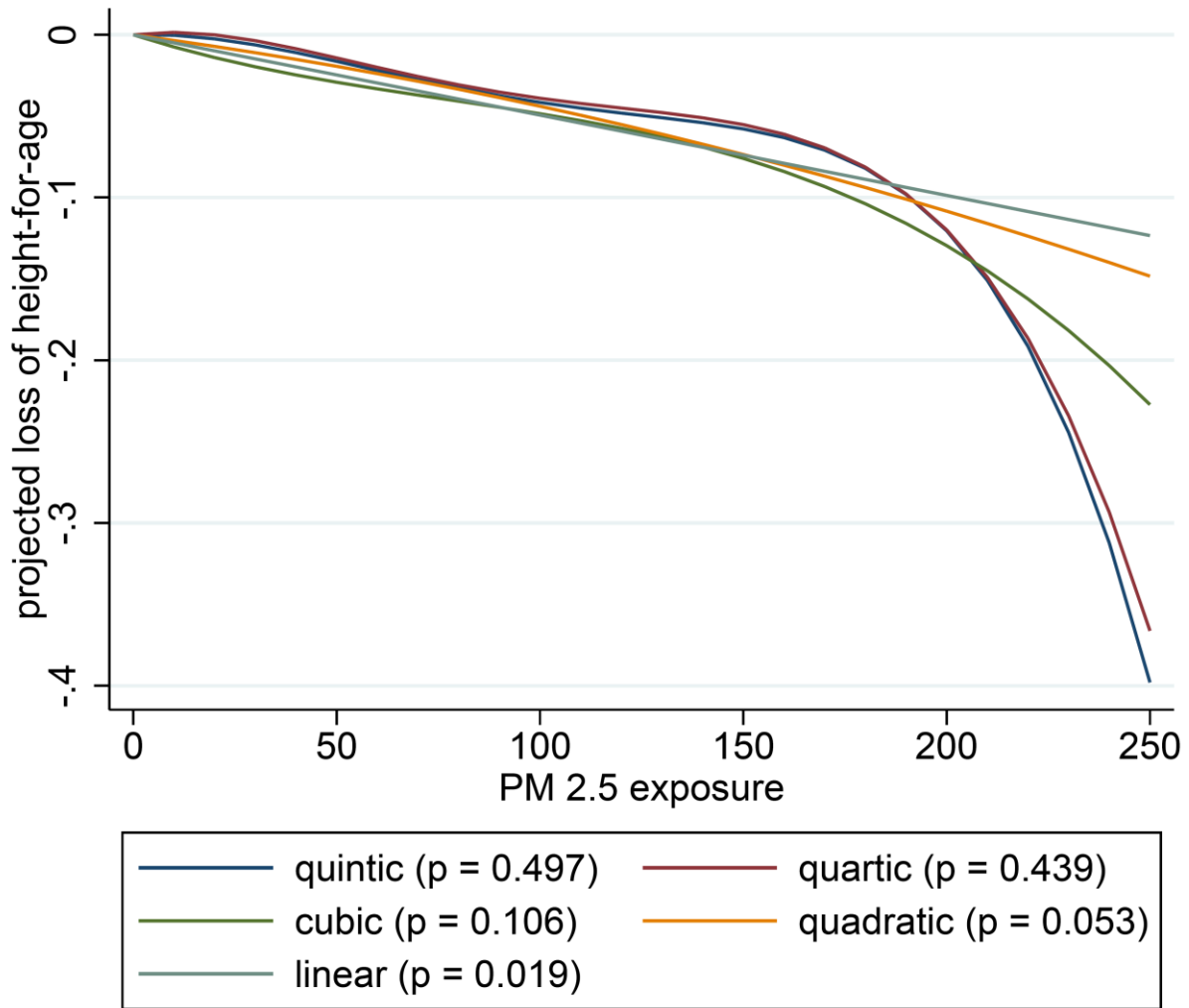
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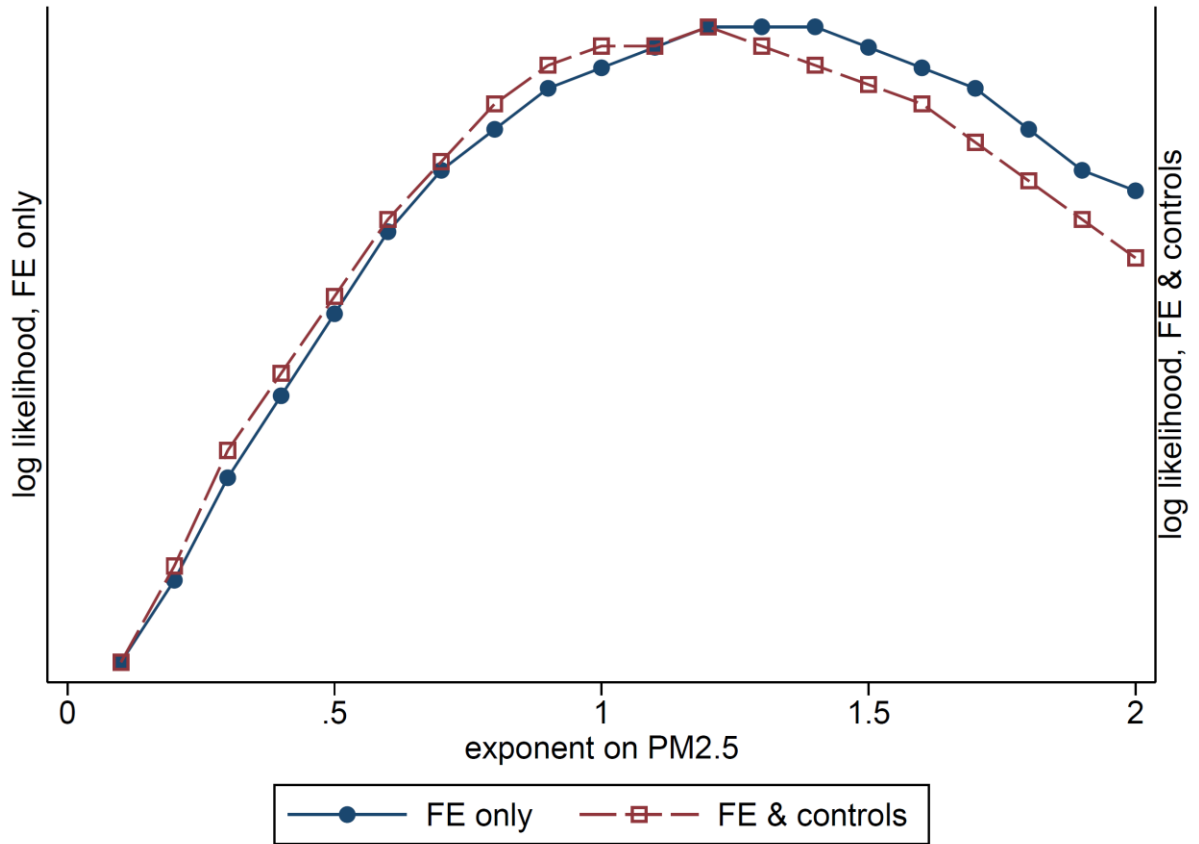
482 **Figure 6.** Projected effects of PM_{2.5} of child height-for-age, at increasing non-linearity. Each
 483 curve is the projected effect from a separate fixed effects regression where PM_{2.5} in the month of
 484 birth is specified as a polynomial of degree 1 through 5. *p*-values report joint *F* tests that all PM_{2.5}
 485 terms are zero.

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490 **Figure 7.** Box-Cox transformation of PM_{2.5} in month of birth: Log likelihood. Each point plots
491 the log likelihood of a separate fixed effects regression of PM_{2.5} transformed according to the
492 coefficient on the horizontal axis.